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Key Points:

- Airborne flux observations for a suite of reactive gases
- Direct constraints on production, loss, and surface-atmosphere exchange
- Comprehensive evaluation of chemical budgets over a temperate forest

Supporting Information:

- Texts S1–S8, Figures S1–S22, and Tables S1–S6

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Quantifying sources and sinks of reactive gases in the lower atmosphere using airborne flux observations

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Abstract Atmospheric composition is governed by the interplay of emissions, chemistry, deposition, and transport. Substantial questions surround each of these processes, especially in forested environments with strong biogenic emissions. Utilizing aircraft observations acquired over a forest in the southeast U.S., we calculate eddy covariance fluxes for a suite of reactive gases and apply the synergistic information derived from this analysis to quantify emission and deposition fluxes, oxidant concentrations, aerosol uptake coefficients, and other key parameters. Evaluation of results against state-of-the-science models and parameterizations provides insight into our current understanding of this system and frames future observational priorities. As a near-direct measurement of fundamental process rates, airborne fluxes offer a new tool to improve biogenic and anthropogenic emissions inventories, photochemical mechanisms, and deposition parameterizations.

1. Introduction

The convective boundary layer (CBL) is the rapidly mixed lowest 1–2 km of the troposphere in direct contact with Earth's surface. Chemical and physical processes within the CBL shape the composition of the lower atmosphere and its relationship with other Earth systems. Emissions of reactive gases, including volatile organic compounds (VOC) and nitrogen oxides ($\text{NO}_x = \text{NO} + \text{NO}_2$), originate from natural and anthropogenic sources at the surface. Photochemistry transforms these gases into secondary products like ozone (O_3) and organic aerosol, impacting global air quality and climate [Fiore et al., 2012]. Dry deposition removes atmospheric pollutants and supplies nitrogen and reactive oxygen to vegetation. Turbulent mixing governs the evolution of CBL structure and entrainment of overlying air masses [de Arellano et al., 2011]. Our ability to predict the state of the lower atmosphere hinges on our comprehension of these processes.

The interplay of biology, meteorology, chemistry, and surface interactions makes the biosphere-atmosphere interface a particularly complex region of the CBL (Figure 1). Forests emit copious VOC including isoprene, a conjugated diene that comprises one third of total global VOC emissions [Guenther et al., 2012]. Oxidation of

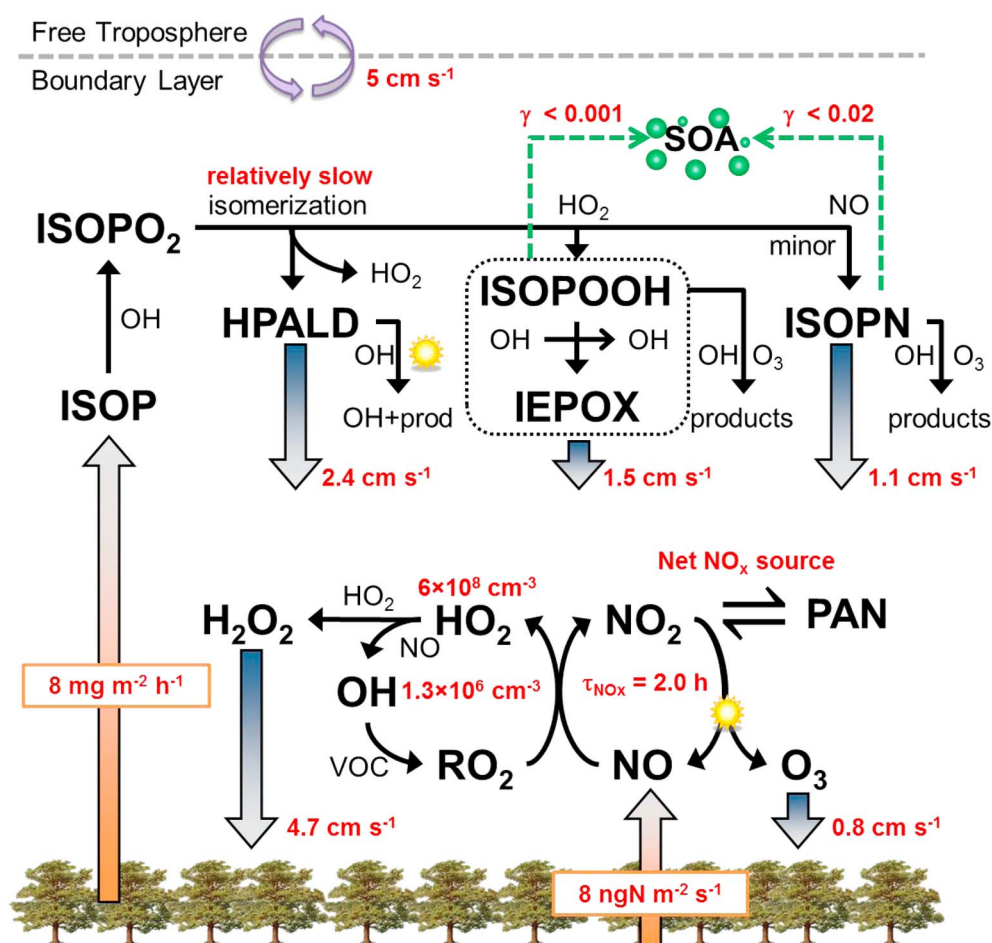


Figure 1. Key processes in the forested daytime CBL. Red text highlights parameters quantified via flux analysis. Orange arrows denote emission, blue arrows denote deposition, and green dashed arrows denote heterogeneous uptake. Only reactions central to the discussion are shown.

isoprene by hydroxyl radical (OH) generates myriad oxygenated VOC (oVOC) that may undergo further reaction, deposit to forest surfaces or partition to secondary organic aerosol (SOA) [Goldstein and Galbally, 2007]. This chemistry also fuels ozone production in the presence of NO_x , which is emitted via combustion and soil microbial processes. Termination of radical cycling produces compounds like hydrogen peroxide (H_2O_2) and nitric acid (HNO_3), which are lost via dry deposition or rainout and may foster feedbacks through impacts on carbon and nitrogen cycles [Magnani et al., 2007].

While our understanding of this system is evolving rapidly, considerable uncertainties persist. Isoprene emission inventories can differ by factors of 2 or more [Hogrefe et al., 2011; Warneke et al., 2010]. Newly discovered isoprene degradation products can be potent radical and aerosol precursors [Paulot et al., 2009; Peeters et al., 2009], yet models still struggle to explain observed radical concentrations [Stone et al., 2012] and SOA mass [Heald et al., 2011]. Most oVOC likely undergo deposition or bidirectional exchange [Nguyen et al., 2015; Park et al., 2013], but measurements are too sparse to adequately constrain these processes. Even in the case of ozone, for which field observations are plentiful, deposition fluxes over forests can vary from model to model by factors of 3 or more [Hardacre et al., 2014].

We present a unique data set that combines the payload and spatial coverage of a flying laboratory with the strengths of eddy covariance (EC) to constrain fundamental CBL processes. The EC technique affords a direct measure of vertical fluxes, which are sensitive to both surface and in situ processes. Previous work, for example, has applied airborne EC methodology to study ozone [Lenschow et al., 1981] and isoprene [Karl et al., 2013; Misztal et al., 2014]. Figure 1 summarizes key parameters derived in our analysis. The

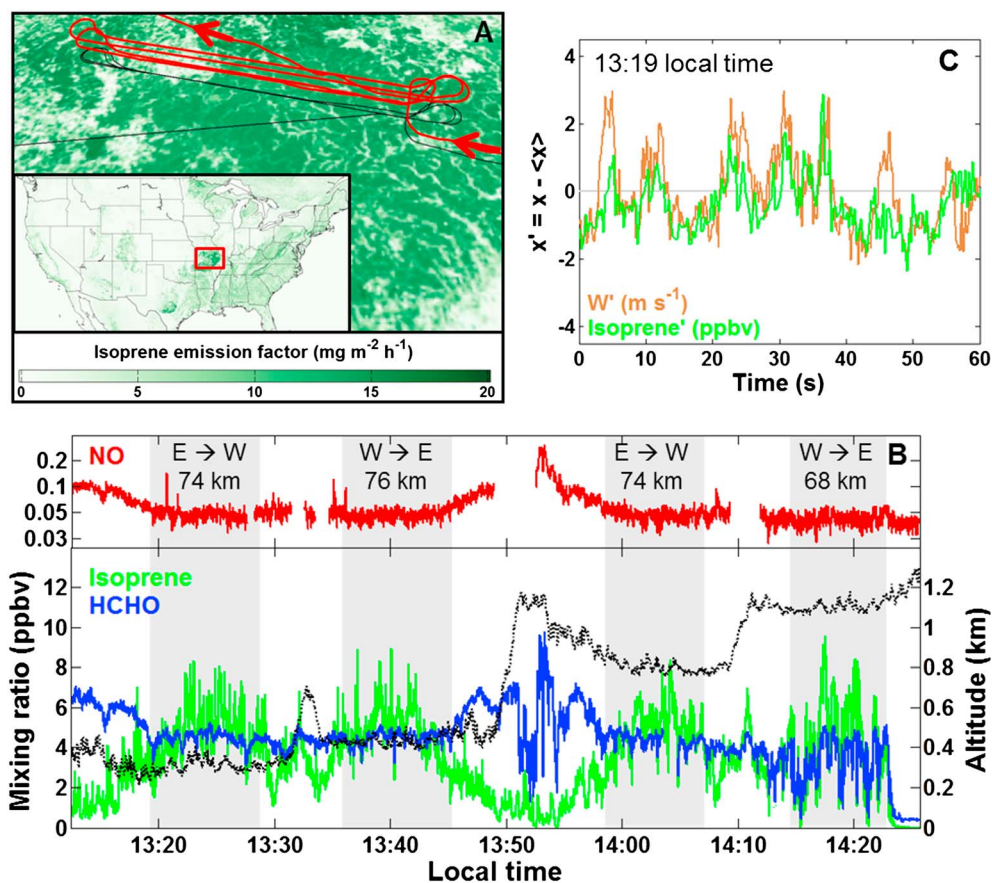


Figure 2. (a) Map of the study area and flight pattern (red line: flight track and black line: ground track). Green shading corresponds to isoprene emission factors from the latest version of MEGAN [Guenther *et al.*, 2012]. The inset shows emission factors for the whole U.S. with a red box indicating the Ozarks. (b) Observed 1 Hz mixing ratios of isoprene, formaldehyde (HCHO) and NO, along with aircraft altitude (broken line) relative to ground level. Shading highlights the data segments used in flux calculations. (c) Example correlation of isoprene and vertical wind speed for a portion of the first flight leg. W' and isoprene' represent the instantaneous deviation from the whole flight leg mean and 20 s running mean, respectively.

synergistic information obtained from simultaneous fluxes of multiple reactive species provides a self-consistent picture of the CBL and a benchmark for testing our understanding of isoprene-oxidant chemistry and biosphere-atmosphere interactions.

2. Mission and Methods

We utilize measurements collected on 6 September 2013 over the Ozark Mountains during the NASA SEAC⁴RS (Studies of Emissions and Atmospheric Composition, Clouds and Climate Coupling by Regional Surveys) mission. Figure 2a maps the flight track together with isoprene emission factors from MEGAN (Model of Emissions of Gases and Aerosols from Nature) [Guenther *et al.*, 2012]. Colloquially known as the “isoprene volcano” [Wiedinmyer *et al.*, 2005], this region is a dense oak forest with sparse local anthropogenic emissions. The NASA DC-8 entered from the east at 13:10 local time and executed a series of four vertically stacked 75 km long transects in the CBL. Weather was mostly sunny and warm (potential temperature = 302 K) with uniform horizontal winds from S-SE at 3 m s^{-1} and a strong subsidence inversion at $z_i = 1060 \text{ m}$. These conditions are near ideal for evaluating turbulent fluxes.

Observations of isoprene, formaldehyde, and nitric oxide (NO) reveal two distinct chemical regimes during this flight (Figure 2b). Elevated NO on the eastern end indicates a pollution plume, likely transported from an upwind power plant. Enhanced NO_x accelerates the radical turnover rate in this air mass, converting isoprene to formaldehyde and other products. In the middle and western portions, NO mixing ratios of 0.05 ppbv are more typical of the continental background, while isoprene is elevated due to both strong

emissions and slower oxidation. Our analysis will focus on the latter region, indicated by the shaded segments in Figure 2b. Section S1, Figures S1–S4, and Table S1 in the supporting information provide further details on the flight location and measurements.

The high variability of isoprene in the CBL illustrates the interplay of chemistry and turbulence. Isoprene is emitted at the surface and transported vertically via turbulent eddies, while OH and ozone destroy isoprene as it mixes. Thus, upward moving eddies are enriched in isoprene compared to downward moving eddies (Figure 2c). This concept forms the core of our methodology: by exploiting the natural variability created by turbulent mixing, we can extract quantitative information on both surface-atmosphere exchange and in situ chemical transformations.

For each of the four flight segments, we calculate EC fluxes of isoprene, hydrogen peroxide (H_2O_2), ozone (O_3), NO_x , peroxyacetyl nitrate (PAN), several isoprene oxidation products, and sensible heat. In addition to traditional EC, we also apply wavelet transforms to obtain horizontally resolved surface fluxes [Misztal *et al.*, 2014]. Details on flux calculations, including spectral and error analysis, are available in Text S2. The flux footprint half-widths for legs 1 to 4 are 0.6, 1.0, 1.4, and 2.1 km, increasing with height (Text S2). Where possible, surface fluxes are compared to output from a $25 \times 25 \text{ km}^2$ GEOS-Chem 3-D global chemical transport model simulation [Wang *et al.*, 1998]. Derived radical concentrations (section 3.4) are compared to both GEOS-Chem and a 0-D box model simulation incorporating the Master Chemical Mechanism (MCMv3.2) [Jenkin *et al.*, 1997; Saunders *et al.*, 2003]. Both models are described in Text S7.

To interpret vertical flux profiles (Figures 3 and 4), we apply the scalar budget equation,

$$\frac{\partial C}{\partial t} = Q - A - \frac{\partial F}{\partial z} \quad (1)$$

which states that the time rate of change in concentration of a scalar, C , is the sum of rates of in situ production and loss (Q), horizontal advection (A , also known as horizontal flux divergence), and vertical flux divergence. Rearranging, integrating over altitude (z) and assuming all terms are approximately constant for a well-mixed CBL over the ~ 1 h flight pattern yields a simplified description of the flux profile:

$$F(z) = \left(Q - A - \frac{\partial C}{\partial t} \right)_{\text{avg}} z + F_0 \quad (2)$$

The intercept (F_0) reflects surface exchange (emission or deposition), while the slope contains the combined effects of chemistry, advection, and storage (the vertical integral of $\partial C/\partial t$). The latter two terms are estimated from concentration data along the transects (Text S3), and the chemical term is calculated as the residual sum of flux divergence, advection, and storage (Table S2). For species where Q varies strongly with altitude (e.g., Figures 4a and 4b), this linear relationship can be replaced with a more rigorous treatment of vertically varying process rates (Text S4). All quoted uncertainties represent 95% confidence intervals propagated from error-weighted least squares fits to flux profiles.

3. Results and Discussion

3.1. Isoprene Emissions

Despite decades of research spanning leaf level to continental scales, accurate parameterization of biogenic VOC emissions remains challenging. Previous observations of in situ isoprene concentrations [Carlton and Baker, 2011] and satellite-derived formaldehyde columns [Millet *et al.*, 2008] indicate that MEGAN overestimates isoprene emissions by 30 to 50% over the Ozarks. Extrapolation of the isoprene flux profile (Figure 3a) yields a surface flux of $8 \pm 1 \text{ mg m}^{-2} \text{ h}^{-1}$, 40% lower than the mean value of $14 \text{ mg m}^{-2} \text{ h}^{-1}$ calculated within GEOS-Chem using MEGAN. Wavelet analysis (Figure 3b) illustrates that local emissions vary by more than a factor of 2 over the transect and that the greatest model-measurement disagreement occurs at the eastern end, near the heart of the Ozarks (Figure S20). Work is currently underway to apply airborne fluxes to a broader evaluation of MEGAN in other regions [Karl *et al.*, 2013; Misztal *et al.*, 2014]. Such measurements will be invaluable for interpretation of future high-resolution satellite observations, linking global-scale continuous emission estimates and ecosystem-level responses.

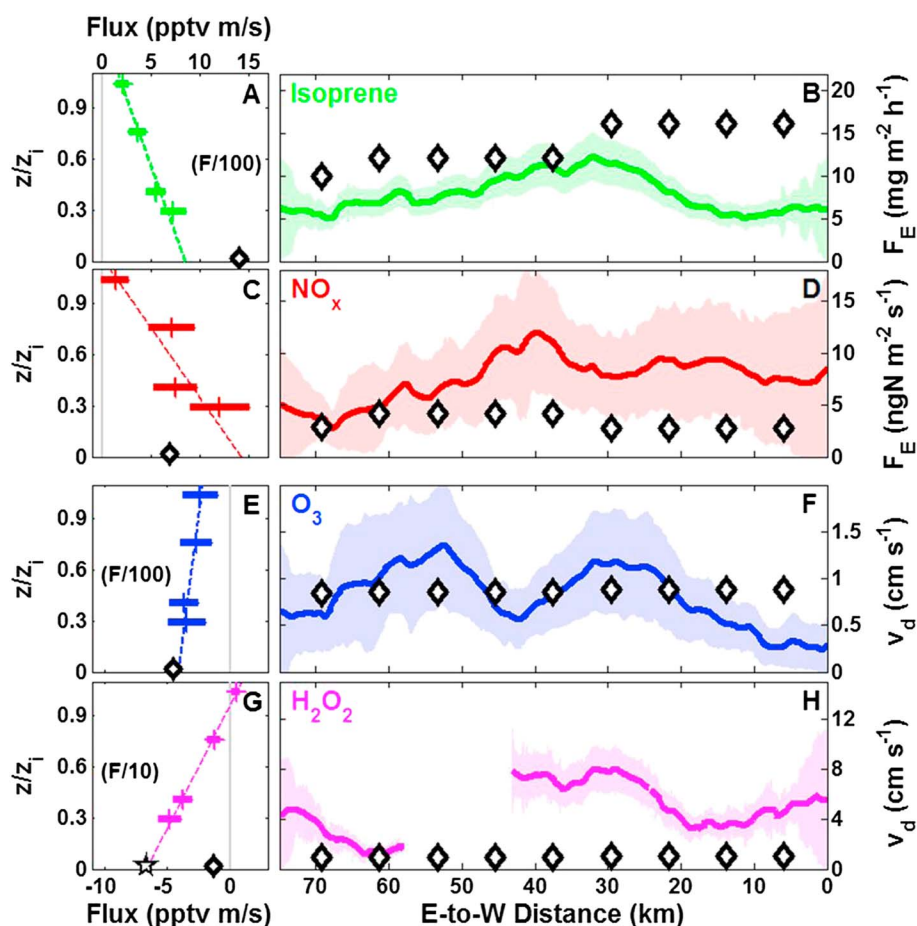


Figure 3. (a, c, e, and g) Vertical flux profiles for isoprene, NO_x , ozone, and H_2O_2 . The top scale corresponds to the upper two plots, and the bottom scale to the lower two. Scaling factors are indicated in parentheses (for left plots only). Altitude is normalized to the CBL height ($z_i = 1060$ m). Crosses and bars represent observed fluxes and 95% confidence intervals. Dashed lines are linear least squares fits. The star in Figure 3g is the theoretical H_2O_2 flux assuming no surface resistance (Text S6). Diamonds represent average surface fluxes from a GEOS-Chem simulation along the flight track. (b, d, f, and h) Horizontal profiles of surface fluxes. Solid lines and shaded areas represent observed fluxes and 95% confidence intervals derived from wavelet transforms along the first transect, corrected to an assumed canopy height of 20 m using observed flux divergence and smoothed with a 10 km running mean (section S2.4 in Text S2). Emission fluxes (F_E) and deposition velocities (v_d) are given in traditional units. Diamonds represent GEOS-Chem results.

3.2. Emissions and Fate of NO_x

Soils emit 15% of global NO_x and are a primary source of reactive nitrogen in many rural areas [Jaeglé *et al.*, 2005]. Airborne fluxes can provide a direct constraint on NO_x emissions at an aggregated scale appropriate for comparison to models and satellite observations. It is most appropriate to consider the total NO_x flux (Figure 3c), as NO and NO_2 interconvert rapidly. The extrapolated surface NO_x emission of $8 \pm 3 \text{ g N m}^{-2} \text{ s}^{-1}$ is substantially higher than the GEOS-Chem flux of $4 \text{ ng N m}^{-2} \text{ s}^{-1}$, which represents the net flux of NO from soils (75%), fertilizer (14%), and anthropogenic sources (18%) less NO_2 deposition (−7%). Modeled NO_x emissions agree reasonably well with observations in the west but are substantially lower in the east (Figure 3d). Model-measurement differences may stem from soil heterogeneity or environmental response functions (e.g., temperature and soil moisture). Accurate representation of soil NO_x emissions is crucial for modeling atmospheric chemistry in pristine forests (e.g., the Amazon) and will be increasingly important in the U.S. as anthropogenic NO_x emissions continue to decline.

Conversion to higher oxides of nitrogen governs the lifetime of NO_x and thus the spatial extent of emission impacts. The combination of flux divergence and storage (Table S2) implies a net chemical

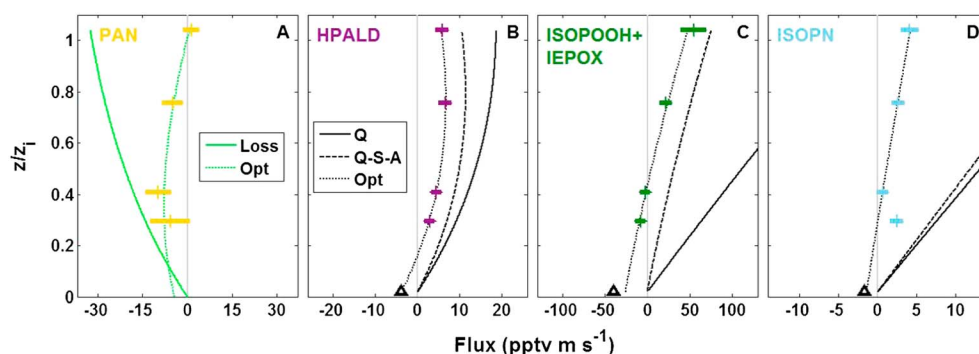


Figure 4. Vertical flux profiles of PAN and isoprene oxidation products. Crosses and bars represent calculated fluxes and 95% confidence intervals. (a) The solid green line is the predicted PAN flux from thermal loss only, while the dotted green line is the “optimized” flux including storage, advection, and primary PA production (section S4.5 in Text S4). (b–d) Solid lines are the “chemical flux” profiles predicted from consideration of gas-phase chemistry (section S4.6 in Text S4). Dashed lines are the sum of chemical, storage, and advection fluxes. Dotted lines are optimized profiles that include the above three terms plus surface deposition and missing production/loss rates (section S4.6 in Text S4). Triangles represent top-of-canopy fluxes calculated from daytime average deposition velocities for a similar forest [Nguyen *et al.*, 2015].

lifetime of 2.0 ± 0.5 h. For comparison, the NO_x lifetime estimated from consideration of formation of nitric acid, alkyl nitrates (ANs), and peroxyacetyl nitrate (PAN) is 1.4 ± 0.4 h (section S4.4 in Text S4). This difference implies an additional in situ NO_x source. Formation ANs comprises $>90\%$ of calculated gross NO_x loss, but subsequent processing of these compounds may release some NO_x . Ascribing the entire difference between observed and calculated NO_x lifetimes to such a mechanism yields a bulk AN recycling efficiency (NO_x produced per NO consumed) of $29 \pm 29\%$, bracketing the range of previous estimates for isoprene-rich environments [Horowitz *et al.*, 2007; Perring *et al.*, 2009]. Our results indicate that NO_x recycling via isoprene-derived ANs is nonnegligible and may be an important NO_x source in remote regions [Paulot *et al.*, 2012].

PAN, which exists in thermal equilibrium with its precursors (peroxyacetyl radical (PA) and NO_2), can be a source or sink of NO_x . Fluxes of PAN are sensitive to temperature [Wolfe *et al.*, 2009], resulting in vertical curvature as loss decreases with altitude (Figure 4a). Analysis of PAN fluxes (section S4.5 in Text S4 and Figure S21) reveals that PAN is a net source of NO_x in this case, with a CBL-average production rate of 0.006 ± 0.003 pptv s^{-1} . For comparison, the soil NO_x source is equivalent to 0.013 ± 0.005 pptv s^{-1} when distributed over the CBL. Thus, transport and decomposition of PAN may account for $\sim 30\%$ of total NO_x production in this region. In the supporting information, we further demonstrate that PAN fluxes can provide a top-down constraint on PA production, which is relevant to both the fate of NO_x and overall VOC oxidation [LaFranchi *et al.*, 2009].

3.3. Deposition of Ozone, H_2O_2 , and oVOC

Dry deposition—the uptake of gases by terrestrial surfaces—is typically described by a deposition velocity, v_d , which intrinsically depends on turbulence, gas, and surface properties. Variants of the resistance framework of Wesely [1989] are widely used to calculate v_d , but observational constraints for these parameterizations are often limited. Extrapolation of airborne fluxes to the surface provides a direct measure of v_d for several species.

Despite a substantial body of laboratory and field work on ozone deposition, global models may overpredict or underpredict measured ozone fluxes by as much as a factor of 2 [Hardacre *et al.*, 2014]. We derive an ozone v_d of 0.8 ± 0.1 cm s^{-1} over the Ozarks, in agreement with the mean GEOS-Chem value of 0.9 cm s^{-1} (Figure 3e). The structure of wavelet fluxes (Figure 3f) reflects a combination of surface heterogeneity and varying chemical conditions. For example, the local minimum near the middle of the transect (Figure 3f) is coincident with the maximum NO_x flux (Figure 3d) and an invariant H_2O_2 deposition velocity (Figure 3h), consistent with enhanced ozone production at this location. This behavior underscores the value of a comprehensive payload and the need to exercise caution when interpreting reactive gas fluxes.

H₂O₂ is an important participant in aqueous-phase chemistry [Ervens *et al.*, 2014] and can be a major precursor of plant-damaging reactive oxygen species [Nguyen *et al.*, 2015]. The measured H₂O₂ v_d of $4.7 \pm 0.4 \text{ cm s}^{-1}$ is significantly faster than the model prediction of 1.0 cm s^{-1} (Figure 3g) but close to the value calculated for a negligible surface resistance ($4.6 \pm 0.4 \text{ cm s}^{-1}$). Variability in wavelet-derived v_d (Figure 3h) roughly tracks underlying leaf area on the western end but not the east (Figure S2), possibly implying changes in other surface characteristics or in the (assumed constant) chemical flux divergence. Despite long-recognized problems with peroxide deposition [Ganzeveld *et al.*, 2006; Hall *et al.*, 1999; Nguyen *et al.*, 2015], models continue to use flawed parameterizations. GEOS-Chem overpredicts H₂O₂ concentrations by 60% for this case, likely because H₂O₂ losses are underestimated. We calculate that deposition comprises 83% of total H₂O₂ loss in this environment.

Figures 4b–4d present fluxes for three major isoprene-derived oVOC: hydroperoxyaldehydes (HPALD), the sum of isoprene hydroxyhydroperoxides (ISOPOOH) and isoprene dihydroxyepoxides (IEPOX), and isoprene hydroxynitrates (ISOPN). Each class of compounds represents a distinct oxidation pathway (Figure 1). Deposition velocities (Figure 1 and Table S5), derived from a rigorous treatment of chemical flux divergence (section S4.6 in Text S4), are comparable to those obtained from tower-based observations over another mixed forest in the southeast U.S. [Nguyen *et al.*, 2015]. Deposition lifetimes range from 12 to 26 h (Table S5), challenging the common practice of assigning a single “physical loss lifetime” to all such species in 0-D box models [Edwards *et al.*, 2013]. oVOC deposition can be a major sink for SOA precursors [Knote *et al.*, 2015] and is a conduit for atmosphere-biosphere transfer of carbon, nitrogen, and reactive oxygen [Nguyen *et al.*, 2015; Park *et al.*, 2013]. A concerted effort is needed to retool deposition parameterizations for more accurate representations of physicochemical driving processes.

3.4. Radical Concentrations and Recycling

For species where the chemical term dominates the budget equation, flux divergence affords a near-direct measure of reaction rates. Karl *et al.* [2013] demonstrated that negative isoprene flux divergence (Figure 3a) reflects loss via reaction with OH. Likewise, positive H₂O₂ flux divergence (Figure 3g) signifies production via self-reaction of hydroperoxyl (HO₂) radicals. Using these relationships (sections S4.1 and S4.2 in Text S4), we derive observationally constrained average concentrations of OH ($(1.3 \pm 0.3) \times 10^6 \text{ cm}^{-3}$) and HO₂ ($(6 \pm 1) \times 10^8 \text{ cm}^{-3}$). It is also theoretically possible to quantify total peroxy radicals (HO₂ and RO₂) from ozone flux divergence, but overwhelming contributions from storage and advection confound this calculation in the present case (section S4.3 in Text S4).

Models routinely underpredict radical concentrations in high-isoprene, low-NO_x regions, but the extent to which this implies problems with measurements or mechanisms remains debated [Mao *et al.*, 2012; Rohrer *et al.*, 2014]. Comparison of flux-derived radical concentrations with GEOS-Chem and a measurement-constrained 0-D box model (Table S6) shows good agreement for HO₂, while OH is underpredicted by 75% and 30%, respectively. Greater underprediction of GEOS-Chem may be partly explained by overestimated isoprene emissions as discussed above. The discrepancy with the box model is comparable to the typical uncertainty of direct OH measurements and smaller than some previously reported model-measurement differences in similar regions [Stone *et al.*, 2012]. Sensitivity simulations performed with the box model reveal that adoption of the latest theoretical HPALD production rate [Peeters *et al.*, 2014] has no impact on model OH (Table S6). Reducing the isoprene + OH rate constant by 50% to mimic reactant segregation [Krol *et al.*, 2000] proportionally raises OH, but reduced radical production is inconsistent with oVOC fluxes (see below). Both modifications degrade HO₂ agreement. The SEAC⁴RS payload did not include HO_x observations. Such a comparison on future missions would facilitate efforts to investigate measurement artifacts in biogenic-rich environments [Fuchs *et al.*, 2011; Mao *et al.*, 2012].

Isomerization of isoprene hydroxyalkylperoxy radicals (ISOPO₂) has emerged as a promising solution for underprediction of HO_x in biogenic regions [Peeters *et al.*, 2009]. This mechanism efficiently recycles radicals and bypasses radical-terminating peroxide formation in low-NO_x conditions. Its impact hinges on the isomerization rate constant, for which the latest theoretical estimate exceeds the experimentally inferred value by a factor of 1.8 [Crouse *et al.*, 2011; Peeters *et al.*, 2014]. HPALDs are a unique and quantitative marker for this chemistry. The temperature dependence of isomerization, which slows by a factor of 2.2 between the ground and 1 km, manifests in the curvature of the HPALD flux profile (Figure 4b). Also shown in Figure 4b is the

theoretical flux profile obtained from consideration of vertically varying production and loss rates in analogy to equation (2) (section S4.6 in Text S4). With the experimental HPALD production rate [Crouse *et al.*, 2011], we are able to close the HPALD flux budget. To our knowledge, this represents the first observationally constrained evaluation of this mechanism.

3.5. Aerosol Uptake

Isoprene oxidation products may comprise ~50% of global SOA [Henze and Seinfeld, 2006]. SOA growth is a complex, multiphase problem, but from the perspective of gas-phase precursors, it is a first-order loss. Flux divergence provides a direct constraint on the net chemical tendency of oVOC (section S4.6 in Text S4), which can be compared with theoretical predictions to quantify potential missing sources or sinks.

IEPOX can be a major contributor to SOA mass in regions with low-NO_x and high sulfate activity [Xu *et al.*, 2015]. The theoretical chemical flux for ISOPOOH + IEPOX exhibits a strong positive slope (Figure 4c) due to fast production of ISOPOOH and slow gas-phase oxidation of IEPOX. Flux divergence is considerably more shallow, but most of this difference reflects a large positive storage term (Table S2). Thus, we can close the budget of ISOPOOH + IEPOX through consideration of gas-phase chemistry alone. These results support an upper limit IEPOX uptake coefficient (γ) of 0.001 ± 0.004 , representing the probability that a gas molecule will be taken up upon contact with an aerosol surface. In contrast, a recently published parameterization [Gaston *et al.*, 2014] gives a significantly faster γ of 0.04 ± 0.03 . Other limiting factors may be at play here, such as suppression of uptake by organic coatings [Gaston *et al.*, 2014] or phase separation [Smith *et al.*, 2012]. Furthermore, the effective Henry's law coefficient for IEPOX is not well constrained. Using a lower value as recommended by McNeill *et al.* [2012] gives a parameterized γ of 0.010 ± 0.006 .

ISOPN may also partition to aerosol, but considerably less attention has been devoted to this process. Observed ISOPN flux divergence is shallower than that predicted with gas-phase chemistry, even after including storage and advection (Figure 4d). This difference implies a significant missing ISOPN sink. Attributing the difference to aerosol uptake yields an upper limit γ of 0.02 ± 0.01 . Recent lab work has revealed that hydrolysis rates vary by more than 2 orders of magnitude between ISOPN isomers [Jacobs *et al.*, 2014]. Our results, which represent a weighted sum for all ISOPN isomers, indicate that heterogeneous loss may comprise 70% of the total ISOPN sink in this environment.

Flux observations provide a valuable constraint on gas-aerosol partitioning, but the subsequent fate of organics remains unclear. In particular, it is unknown what fraction of the organic carbon evaporates following heterogeneous processing. Simultaneous measurements of oVOC fluxes and aerosol composition will provide better closure of the reactive carbon budget and a more complete understanding of the isoprene-SOA system.

3.6. Entrainment

Fluxes of short-lived compounds like isoprene are well suited for quantifying the rate of mixing between the CBL and the overlying free troposphere (FT) [Karl *et al.*, 2013]. The entrainment flux can be parameterized as $v_e \Delta C$, where v_e is an entrainment velocity and ΔC is the CBL-FT concentration difference. Effectively a measure of boundary layer growth, v_e will vary in both time and space. Isoprene is well suited for this calculation due to its large CBL-FT concentration gradient. Using observed isoprene fluxes and concentrations, we calculate $v_e = 5 \pm 1 \text{ cm s}^{-1}$ for this well-developed CBL, within the range of values reported by Karl *et al.* [2013]. Figure S22 compares entrainment velocities calculated from other flux profiles; those that do not agree with the isoprene-derived v_e typically exhibit weak CBL-FT concentration gradients or significant storage terms. This parameter is rarely constrained by observations but can be crucial for representing the diurnal evolution of boundary layer composition [de Arellano *et al.*, 2011].

4. Outlook

The ecosystem-level view inherent to airborne fluxes is ideal for refinement of model parameterizations and mechanisms. Future airborne missions could acquire fluxes over diverse surfaces (forest, farmland, ocean, urban and industrial areas, etc.) and under contrasting environmental conditions to build an extensive data set. Such measurements can be obtained in tandem to other mission objectives, adding value to a chemistry-oriented payload. The EC technique inherently requires turbulence and is thus best applied in unstable

daytime boundary layers. We stress that the synergistic information afforded by a broad instrument suite is crucial for developing a self-consistent picture of the CBL.

Airborne fluxes hold promise well beyond the cases presented here. For example, simultaneous measurements of speciated and total oxidized nitrogen (NO_y) would permit characterization of the NO_y flux budget and identification of unknown processes, such as nitrous acid production [Li *et al.*, 2014]. Other potential topics include the link between isoprene emissions and plant productivity [Unger *et al.*, 2013], bidirectional ocean exchange of small-chain α VOC [Coburn *et al.*, 2014], and aerosol chemistry and composition [Farmer *et al.*, 2013].

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